Sorption Of Rare Metal lons by Ionite Based on Diglycidyl Thiourea and Various Amines

F.B. Eshkurbonov¹, E.R. Safarova^{1*}, B.Y. Ismoilov²

¹ Department of Chemical Technology. Termez Institute of Engineering and Technology

² Termiz Branch of Tashkent Medical Academy

Abstracts: The article studies the process of sorption of molybdenum and vanadium ions on anion exchangers obtained on the basis of thiourea, epichlorohydrin and various amines. Various amines (urea, guanidine, melamine, polyethylene polyamine, ethylenediamine, etc.) are used as substances containing ionic groups. The regularities of ion exchange and the main properties of the tested anionites are given, the influence of the pH of the medium and interfering ions in the process of sorption of molybdenum and vanadium ions are studied. For vanadium, the use of empirical partition relations and of an extended Freundlich equation were tested, with promising results. The objective of this paper is to investigate the use of the pH dependent Freundlich model of Gustafsson et al., for describing vanadate(V) sorption to 26 mineral soils (all having less than 12% organic C), and to discuss the possible use of the model for risk assessments. The kinetic curves of the sorption process, the values of the diffusion coefficients for the initial periods of molybdenum sorption are given. In particular, the distribution coefficient of the radiotracers of ⁹⁹Mo and ¹³¹I as homologs of Sg and Ts using the surface modification of chabazite was investigated as liquid-solid system in the present study. The kinetics of molybdenum sorption was carried out under static conditions from an ammonium molybdate solution with a concentration of 1 g/L for molybdenum ions and at pH 4.5-5. Under the same conditions, molybdenum ions are sorbed much faster by the DGT + M anion exchanger in comparison with AN-2F anion exchangers, in which the diffusion coefficient is much lower. The static exchange capacity (SEC) was determined in industrial solutions in the presence of mineral acids (hydrochloric, sulfuric and nitric), usually contained in them. The obtained kinetic equilibrium parameters of the tested anion exchangers were compared with those of industrial polycondensation anion exchangers, such as AN-2F and AN-1.

Keywords: Ion Exchanger, Sorption, Molybdenum, Vanadium, Radiotracers, Chabazite, Kinetic Curves, Process, Concentration, Diffusion.

1. INTRODUCTION

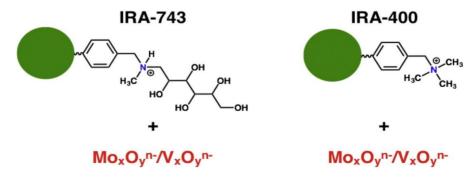
Industrialization and technological developments have important roles in the daily life of people as well as causing of environmental problems. Presence of the metal ions in the environment is very useful for human beings and living organisms while they cause a health problem when high concentrations of these metals directly or indirectly contact with humans, animals, underground water sources and agricultural areas.

Molybdenum is considered as a key micronutrient for some living organisms and one of the biologically active transition element due to its vital role in enzymatic redox reactions as well as its specific geochemical behavior [1]. It is a bioessential element and the presence of molybdenum in water samples is of great interest from chemical and environmental points of view [2]. The decay product of Mo-99, Tc-99m, is the workhorse isotope in nuclear medicine for diagnostic imaging. Tc-99m is used for the detection of disease and for the study of organ structure and function. Tc-99m is especially useful for nuclear medicine procedures because it can be chemically incorporated into small molecule ligands and proteins that concentrate in specific organs or tissues when injected into the body. The fast-growing utilization of molybdenum in the anthropogenic activities illustrates the potential for a lot of toxic discharges into the ecosystem. Molybdenum commonly presents principally in the hexavalent oxidation state, Mo(VI) and molybdate (MoO_4^{-2}) ions [3]. Molybdate oxoanions are water soluble and cause an environmental problem if their concentration exceeds 5 ppm [4]. Pollution of groundwater by molybdate anions represents a serious issue in the field of drinking water obtained from wells [5].

Excited and advanced studies on nuclear chemistry are directed to discover and identify the chemical properties of superheavy elements (SHEs). All elements of atomic number \geq 104 are called SHEs [6]. SHEs are synthesized by accelerators in heavy-ion-induced nuclear reactions [7]. The chemical studies of SHEs are big challenged due to 709

SHEs experiments must be chemically performed on a "one-atom-at-atom" scale because of extremely low production rates of atoms and the shortlived (few seconds) of these nuclides. Preparation of these kinds of experiments is required a lot of expense and preliminary investigations. Therefore, the development of new chemical systems applicable in fast kinetic experiments usually the lighter homologs was used as model experiments of SHEs [6–10]. Seaborgium (Sg) and tennessine (Ts) are located in the periodic table in groups 6 and 17 as Mo and I, respectively. In our previous work, the extraction and sorption behavior of carrier-free ⁹⁹Mo as a homolog of Sg using ionic liquids [6] and impregnated resin with trioctylamine (BR-TOA) [7] were investigated. In this regards, the distribution coefficient of the radiotracers of ⁹⁹Mo and ¹³¹I as homologs of Sg and Ts using the surface modification of chabazite was investigated as liquid-solid system in the present study. The largest producers of molybdenum separate the isolated metal from copper compounds using selective flotation [11].

Vanadium is a transition metal that, being ubiquitously distributed in soil, crude oil, water and air, also found roles in biological systems and is an essential element in most living beings. There are also several groups of organisms which accumulate vanadium, employing it in their biological processes. Vanadium being a biological relevant element, it is not surprising that many vanadium based therapeutic drugs have been proposed for the treatment of several types of diseases. Namely, vanadium compounds, in particular organic derivatives, have been proposed for the treatment of diabetes, of cancer and of diseases caused by parasites. The recovery of vanadium is more complicated and employs a vanadium-bearing slug from steel manufacturing or the removal of extremely small vanadium levels from copper-containing concentrates [12]. In both cases, multiple solvent extraction methods are been applied [13,11]. However, when the concentrations of the materials that are being recovered are low, the approach is not sufficient. Moreover, applying these procedures carries the risk of losing the precious resources in the substantial levels of wastes that are also produced [11]. According to Syed [14], in this a situation, the only economically viable method for achieving a satisfactory extraction degree is the application of ion-exchange sorbents bearing selective ligands. Over the last several years, multiple polymeric materials are been applied for the sorption and recovery of molybdenum and vanadium. The successful application of modified poly(ethyleneimine) supports containing ammonium, carboxymethyl or phosphonomethyl groups [15], iminodiacetic chelating resins [16] and D2EHPA-immobilized Amberlite XAD-4 resin [17] have been reported. Further applications indicate that commercially available ion-exchange suspension copolymers, such as Amberlite IRA904 [18] and Dowex 1-x8 [19], are suitable for preconcentration of molybdenum and vanadium. In addition, modified silica polyamine composites (SPC) made from silanized amorphous nanoporous silica gel and polyamines have been employed for the selective oxyanion removal to achieve a higher selectivity towards tungsten than molybdate [20]. Styrene/divinylbenzenebased resins (i.e., Amberlite IRA-400 and Amberlite IRA-743 (see Fig 1)) bearing guaternary ammonium and Nmethyl-D-glucamine functionalities exhibit the ability to effectively retain oxyanions, such as borates [21,22], chromates [23] and arsenic [24,25]. The current study aims to investigate the suitability of Amberlite IRA400 and Amberlite IRA-743 resins for sorption of molybdenum and vanadium oxyanions from their single-component aqueous solutions.





The sorption kinetics as well as the environment pH at which the polymers exhibited the best performance were investigated. In addition, equilibrium studies were performed different temperatures. Finally, the reusability of the ion-exchange resins was investigated for repeated sorption and desorption cycles.

The iron oxide nanoparticles are increasingly applied for heavy metals removal from wastewaters since they exhibit excellent adsorption properties and can be easily separated using magnetic field [25-27]. Their drawbacks like small particle size, excessive pressure drops and coaggregation observed in flow-through systems could be avoided by supporting magnetite nanoparticles on polymers. Magnetic polymer microspheres combining a polymer and inorganic magnetic nanoparticles have been successfully used as carriers for enzyme immobilization, protein purification [28], separation of toxic and radioactive pollutants [29], etc. Glycidyl methacrylate (GMA) based copolymers have versatile applications due to the presence of epoxy groups which offers numerous functionalization possibilities in mild reaction conditions. Amino-functionalized macroporous crosslinked copolymers of GMA and ethylene glycol dimethacrylate (EGDMA), PGME, prepared by suspension copolymerization in the shape of regular beads and specific pore size [30] have been utilized as matrix for enzyme attachment [31], sorbents for removal of textile dyes [32], radionuclides [33], precious and heavy metals, etc. [34-36].

In industry, when processing non-ferrous metals and their satellites, when selective processes are required in order to concentrate or separate the target components, anion exchange is especially effective if the metal ions are in the form of anionic complexes. Among these metals, an important place is occupied by molybdenum, rhenium, tungsten, vanadium, etc. [37]. In the literature there are a significant number of publications devoted to the study of the features of separation, separation of molybdenum and its purification from impurities using anionites of various structures. [38].

The polycondensation reaction opens up wide possibilities for the synthesis of ion-exchange polymers. When obtaining anion-exchange polymers of this type, the process of formation of the spatial structure and the introduction of ionogenic groups proceed in one stage. Various amines (urea, guanidine, melamine, polyethylene polyamine, ethylenediamine, etc.) are used as substances containing ionic groups. [39,40].

There are different approaches to the development of models for simulation of metal binding in soils. One is to use process-oriented surface complexation models [41,42,43]. However, although these models provide insights to the mechanisms involved, they are difficult to set up for soils due to the large number of parameters that need to be estimated. Another approach is to use empirical isotherm equations. Although they are not process-oriented they can still be useful for predicting anion sorption as they require less input. For vanadium, the use of empirical partition relations [44] and of an extended Freundlich equation [45] were tested, with promising results. The Freundlich model can also be extended with a term accounting for the influence of pH, which is often the most important parameter that governs metal sorption in soils [46]. We recently showed a new way to include such a pH term for describing SO₄-² adsorption to soils [47]. An important prerequisite for the use of a Freundlich model is that only one type of adsorption process is important. In the case of vanadium, the model is not likely to perform well if both vanadium(IV) complexation to organic matter and vanadate(V) sorption to Fe and AI hydrous oxides are significant. Therefore it may be hypothesized that the Freundlich models may work in mineral soils in which vanadate(V) sorption may be the predominant process. The objective of this paper is to investigate the use of the pH dependent Freundlich model of Gustafsson et al. [47] for describing vanadate(V) sorption to 26 mineral soils (all having less than 12% organic C), and to discuss the possible use of the model for risk assessments. To investigate the hypothesis that adsorbed vanadate(V) was the predominant reaction product, vanadium Kedge XANES spectroscopy was performed for three of the soils, which were different concerning pH, content of Fe and AI hydrous oxides, and texture.

2. MATERIALS AND METHODS

At present, it is important to study the physicochemical and mechanical properties of ion-exchange polymers, which make it possible to determine the fields of application, advantages and disadvantages of the synthesized anionites. On the other hand, physicochemical studies of ion exchangers make it possible to determine ways of modifying some properties of anionites.

Despite a significant number of studies devoted to the ion-exchange method for the extraction and separation of metals, solving this problem continues to be a top priority for the hydrometallurgical industry. In the Republic of

Uzbekistan, the number of branches of the national economy that use ion-exchange polymers is increasing annually. Until now, ion exchangers are imported into Uzbekistan from abroad, the use of which affects the cost of products. In addition, most of the imported ion exchangers, especially of the polycondensation type, have low indicators of such properties as chemical heat resistance, mechanical strength, and others, which limits the possibilities and scope of their application.

For this purpose, we carried out a study of the physicochemical and mechanical properties of the obtained ionexchange polymers.

In this regard, it was of interest to study the sorption of molybdenum on the obtained by us anion exchanger DGT+PEPA (diglycidylthiocarbamide and polyethylenepolyamine) DGT+M (diglycidylthiocarbamide and melamine), DGT+HPAN (diglycidylthiocarbamide and hydrolyzed polyacrylonitrile) which contains polycrystalline groups in its structure. [48].

The study of the kinetics of the sorption process makes it possible to reveal the mechanism of sorption, which, in turn, will make it possible to choose the optimal parameters for conducting the ion-exchange process. The kinetics of molybdenum sorption by the anion exchanger DGT+PEPA, DGT+M, DGT+HPAN and the industrial anion exchanger AN-2F was carried out under static conditions from an ammonium molybdate solution with a concentration of 1 g/L for molybdenum ions and at pH 4.5–5. Anionites were used in Cl^- - form.

The contact time of the anion exchangers with the solution ranged from 5 minutes to 2 days. The results of sorption of molybdenum for a time from 0.5 to 2 days remain practically the same. In this case, the static exchange capacity (SEC) of the molybdate ion for 0.5 days on the DGT + PEPA anion exchanger was 282 mg/g, DGT + M 275 mg/g, DGT + HPAN 278 mg/g, on the AN-2F anion exchanger, 262mg/g. To quantitatively characterize the equilibrium distribution of molybdenum between the phases of the anionite and the solution, the distribution coefficients K_d were calculated using the formula [49].

$$K_d = \frac{M_{ex}}{M_s}$$
(1)

Mex - concentration of molybdenum in anion exchanger, mg/l;

Mor - equilibrium concentration of molybdenum in solution, mg/ml;

3. RESULTS

The distribution coefficient for the DGT + M anion exchanger was 413 mg/l, and for the AN-2F anion exchanger, 389 mg/l. The data obtained indicate that the anion exchangers DGT + PEPA, DGT + M, and DGT + HPAN have a higher sorption capacity in relation to molybdenum ions as compared to the AN-2F anion exchanger. The distribution coefficients clearly characterize the sorption and selectivity of anionites. This value is a measure of the affinity of the tested anionites for molybdenum and shows the degree of selectivity under specific research conditions.

In order to establish the mechanism of sorption of molybdenum, we used the known time dependences of the ion exchange rate $lg(1 - F) = K_1\tau$ for film and $F = K_2\tau^{1/2}$ for gel kinetics, where F is the relative content of sorbed molybdenum in the anion exchanger (the degree to which ion-exchange equilibrium is reached) [50]. The study was carried out under static conditions with the ratio of the liquid phase of the ammonium molybdate solution to the solid phase - anionite 1000: 1.

The F value was determined by the equation:

$$F = \frac{M_{\tau}}{M_{\infty}} \qquad (2)$$

 M_{τ} - sorbed molybdenum for time τ , mg.

M_{∞} - sorbed molybdenum at equilibrium

Based on the experimental and calculated data, kinetic curves were constructed (Fig. 2-4) and table. 1.

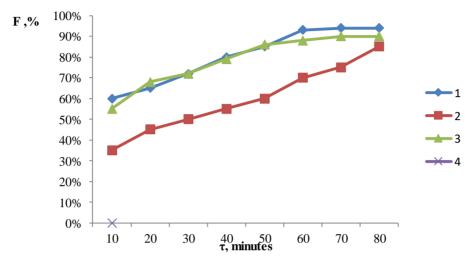


Fig. 2. Molybdenum absorption rate: 1 - anion exchanger DGT + M; 2 - anion exchanger DGT + PEPA; 3 - anion exchanger DGT + HPAN; 4 - anion exchanger AN-2F.

When constructing a graph from the data obtained, it was found that for the dependence of $\lg (I - F)$ on τ , the experimental points do not fit on a straight line. Apparently, the internal diffusion mechanism of kinetics has a dominant effect on the sorption rate.

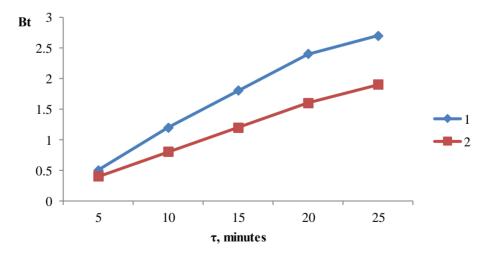


Fig. 3. Graph of diffusion of sorption of ions: 1 - anion exchanger DGT+M; 2 - anion exchanger DGT + PEPA; 3 - anion exchanger DGT + HPAN; 4 - anion exchanger AN-2F.

The limiting effect of gel kinetics can also be judged by the linear dependence in the coordinates $F - \sqrt{\tau}$, when F<0.4 (Fig. 3) and according to the dependence curves $B_t - \tau$ for the whole process, where, in the first approximation, all points fit on a straight line (Fig. 3.).

The diffusion coefficient was determined by the formula:

$$\overline{D} = \frac{B_t \cdot r^2}{\tau \cdot n^2} \qquad (3)$$

Where: B_t - dimensionless quantity, which we find from tabular data as a function of F.

 \overline{D} - diffusion coefficient, cm²/sec;

au - contact time of solution with anionites, sec .;

r - radius of anionite grain in a swollen state, cm;

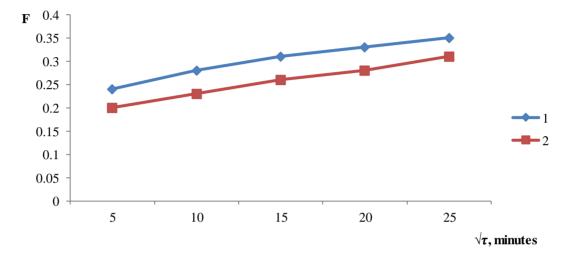


Fig. 4. Graph of diffusion absorption of molybdenum ions: 1 - anion exchanger DGT + M; 2-anion exchanger DGT + PEPA; 3-anion exchanger DGT + HPAN; 4-anion exchanger AN-2F.

The calculated values of the diffusion coefficients of molybdenum (Fig. 4.) after 1 hour of contact of the ion exchanger with the solution are for DGT + M anion exchangers $-5.7 \cdot 10^{-3}$ cm²/sec and for AN-2F $-3.9 \cdot 10^{-3}$ cm²/sec.

The table shows the values of the diffusion coefficients \overline{D} for the initial periods of sorption of molybdenum, when the degree of saturation F of the anion exchanger with molybdenum is not more than 0.5.

Anionite	F	$\overline{D}\cdot 10^8$, ${ m cm}^2/{ m sec}$	au, min	
DGT + M	0,09 0,10 0,19 0,25 0,31 0,34	0,64 0,57 0,85 0,77 0,74 0,67	5 10 15 35 50 60	
AN-2F 0,23 0,37		0,85 0,83 0,72	30 60 120	

 Table 1. Kinetics of sorption of molybdenum on anion exchangers

The data in the table show that at the initial moments of sorption, the diffusion coefficients practically do not remain constant. Despite the fact that the anion exchanger DGT + M surpasses the industrial anion exchanger AN-2F in sorption rate, the kinetics of molybdenum sorption for the tested anion exchangers as a whole is slow, the sorption process is limited by internal diffusion. The slow rate of sorption of molybdenum is apparently explained by its ionic state and steric factors.

Under the same conditions, molybdenum ions are sorbed much faster by the DGT + M anion exchanger in comparison with AN-2F anion exchangers, in which the diffusion coefficient is much lower.

For weakly basic anion exchangers obtained by the polycondensation of melamine, epichlorohydrin and thiourea - (DGT + M), guanidine, epichlorohydrin and thiourea - (DGT + HPAN), studies were carried out of individual regularities of ion exchange, which could serve as the basis for the physicochemical characteristics of the tested anionites. [48].

Of the main chemical properties of ion exchangers, the ion exchange capacity, which characterizes ion exchangers in order to assess their operational properties, is of great practical importance [51]. Its value mainly depends on the number of ionogenic groups of the ion exchanger, the degree of their dissociation, as well as on the nature and concentration of the exchanged ions.

The static exchange capacity (SEC) was determined in industrial solutions in the presence of mineral acids (hydrochloric, sulfuric and nitric), usually contained in them. The obtained kinetic equilibrium parameters of the tested anion exchangers were compared with those of industrial polycondensation anion exchangers, such as AN-2F and AN-1 [52]. Table 2. The main physical and chemical properties of the tested anionites are presented.

ſ		Bulk	Mechanical strength,%	Specific volume of the swollen anionite in the OH ⁻ form, ml/g	SEC 0.1 N solution, mg-eq/l		
	Anionites	weight, g/ml			H_2SO_4	HNO 3	HCI
	DGT + M	0,65	99,1	2,5	7-7,5	4,6	9,5
	AN-2F	0,5	-	3,1	6	-	6,2

Table 2. Basic properties of the tested anionites

The studies were carried out depending on the ionic form of the anionite, the pH of the medium, the presence of competing ions, etc. For comparison, we used the industrial anionite AN – 2F, which is recommended for the extraction of molybdenum from industrial solutions, and the anionite FA –C, which is selective to molybdenum ions [53]. The sorption of molybdenum was carried out under static conditions, the anionites were tested in OH^- , SO_4^- , Cl^- - forms, the initial solutions of ammonium molybdate had a pH in the range from 4 to 5. The sorption of molybdenum was carried out both from pure solutions of ammonium molybdate ($C_{out}=1g/l$) and in the presence of competing sulfate ions ($C_{Na_2SO_4} = 1N$). Table 3. shows the data on the sorption of molybdenum by anion exchangers.

Anionites		Absorbed molybdenum from pure solutions, mg/g		Absorbed molybdenum in the presence of 1 n solution Na ₂ SO ₄ , mg/g
DGT+M				
	in OH^- - form			168
	in SO_4^- - form	245	374	202
	in \mathbf{Cl}^{-} - form	230		175
FA-C				
	in OH^- - form			60
	in ${\rm SO_4}^-$ - form	115 68	212	60 60
	in Cl^- - form			
AN-2F				
	in Cl^- - form	230	320	62

Comparison of data on the sorption of molybdenum on the tested anion exchangers in OH^- , SO_4^{2-} , Cl^- - forms shows that in the cases OH^- , Cl^- - forms, the sorption of molybdenum is somewhat suppressed by sulfate ions, while in SO_4^{2-} - form this phenomenon is absent.

Analysis of the table. 3. shows that the anion exchanger DGT + M has a high selectivity to molybdenum, which is almost the same during sorption, both from pure solutions of ammonium molybdate and in the presence of competing ions. Anionite FA-S is somewhat inferior to the anionite AN-2F in sorption of molybdenum from pure solutions, however, in the presence of competing ions, it has the same sorption capacity.

The influence of the pH of the medium on the sorption of molybdenum was investigated in the range of pH = 1-10. The specified pH value was reached by adding sulfuric acid and ammonium hydroxide. Table 4. presents the results, where the data for the AN-2F anionite are given for comparison.

Anionites	pH of the medium	Sorption of molybdenum from ammonium molybdate solution (C = 1 g / I),%
DGT+M in OH^- - form	2,2 3,8 5,1 8,2 9,6	72 88 74 38 25
AN-2F in OH^- - form	2,2 4,2 6,1 8,1 10	38 54 39 20 8,0

Table 4. Sorption of molybdenum by anion exchangers and dependence on the pH of the medium

From the data in Table 4. it can be seen that the maximum value of molybdenum sorption is observed at pH 4– 6.5. From a practical point of view, an important indicator of ion-exchange polymers is the rate at which equilibrium is reached - the kinetics of sorption.

When choosing an ion exchanger for the sorption extraction of any metal, along with other factors, its kinetic properties are of great importance. In relation to vanadium, in the literature there is a delayed kinetics of sorption by strongly basic anion exchangers in salt form [54]. Therefore, it is of interest to obtain data on the rate of vanadium sorption by the named anionites DGT+PEPA, DGT+M, and DGT+HPAN in the hydroxyl form.

The main task of the work was to elucidate the limiting stage that determines the vanadium absorption rate. These stages can be external diffusion, internal diffusion and ion exchange (chemical) interaction. The results of sorption work are shown in Figure 5. In coordinates, the content of vanadium in the resin is time.

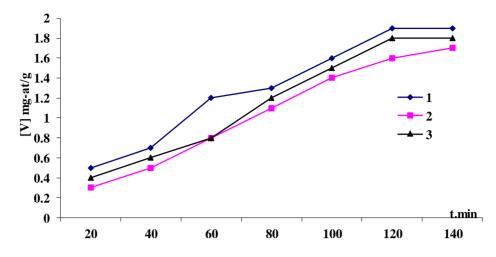


Fig. 5. Dependence of vanadium sorption by the anion exchangers DGT+PEPA(1), DGT+M (2) and DGT + HPAN (3) in the hydroxyl form on the duration of stirring.

From the figure it can be concluded that the rate of vanadium sorption by these anion exchangers decreases in the order DGT+PEPA> DGT+HPAN.

Calculation of the coefficients of the functions that determine the external and internal diffusion shows that the value of the internal diffusion rate coefficient is less than the value of the external diffusion coefficient (Table 5.). Hence, a preliminary conclusion can be drawn that the rate of absorption of vanadium by anion exchangers has a large internal diffusion.

Anionites	External diffusion rate coefficients K·10 ² sec ⁻¹	Internal diffusion rate coefficients B·10 ² sec ⁻¹	
DGT+PEPA	1,75±0,46	4,85±1,7	
DGT+M	1,79±0,35	5,13±1,45	
DGT+HPAN	1,56±0,58	3,92±1,03	

Table 5. The coefficients of the rates of external and internal diffusion, calculated from the experimental data.

In addition to these coefficients, the determining factor can be assessed using kinetic curves in coordinates $\ln(1-F) = \varphi(t)$ and $F = \varphi(\sqrt{t})$. The curves plotted based on the results of the work are shown in Figures 6 and 7.

Figure 6 shows that the kinetics of vanadium sorption by ion exchangers DGT+PEPA, DGT+M and DGT+G in the hydroxyl form is not determined by either external diffusion or chemical interaction, since the curves are not straight lines.

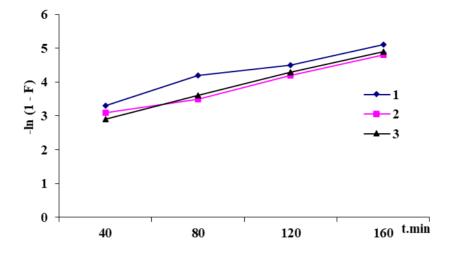


Fig 6. Kinetic curves of vanadium sorption by anion exchangers DGT + PEPA (1), DGT + M (2) and DGT + HPAN (3) in the coordinates of the equation $\ln (1 - F) = \phi(t)$

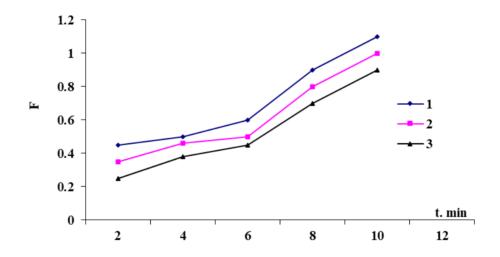


Fig 7. Kinetic curves of vanadium sorption by anion exchangers DGT + PEPA (1), DGT + M (2) and DGT + HPAN (3) in the coordinates of the equation $F = \varphi(\sqrt{t})$

Figure 7 shows the kinetics of vanadium sorption by the anion exchangers DGT + PEPA, DGT + M, and DGT + HPAN. In the case of gel kinetics (internal diffusion), the dependence of F on \sqrt{t} should be expressed by a curve that, for small values of t, has a straight course and then bends. This is the case in our case (Figure 7), which confirms the preliminary conclusion drawn on the basis of table number 5.

The vanadium sorption process is described by the diffusion equation for a spherical shape:

$$\frac{dqr}{dt} = D \frac{d^2 qr}{dx^2} \qquad (4)$$

If the value of the diffusion coefficient D, calculated for different times of contact of the ion exchanger with the solution, is constant, then it can be argued that the rate of absorption of ions is indeed limited by internal diffusion.

To calculate the diffusion coefficient, you can use the ready-made tabular data of the dependence B_t of F and the equation:

$$D = \frac{B_t \cdot r^2}{t \cdot \pi^2} \tag{5}$$

We calculated the diffusion coefficients of vanadium in the anion exchangers DGT + PEPA, DGT + M and DGT + HPAN in the hydroxyl form, which are, respectively, 1,71·10⁻⁶; 1,42·10⁻⁶; and 9,66 ·10⁻⁸; cm²/sec. Using the found diffusion coefficients for the named anion exchangers with a particle size of 0.25 - 0.5 mm, the maximum time required to establish equilibrium between the solution and the ion exchanger under these conditions was calculated. It is 7.24 hours for DGT+PEPA ion exchanger, 8.42 hours for DGT+M ion exchanger and 8.8 hours for DGT+HPAN.

To carry out experiments on desorption of vanadium, depending on the mixing time, a solution of 0.2 n sodium hydroxide was used, considering that this concentration is sufficient for desorption of vanadium from the anionites DGT+PEPA, DGT+M and DGT+HPAN. The experiments were carried out under the same conditions as in the study of the sorption rates. The results of experiments and calculations are shown in Figure 8.

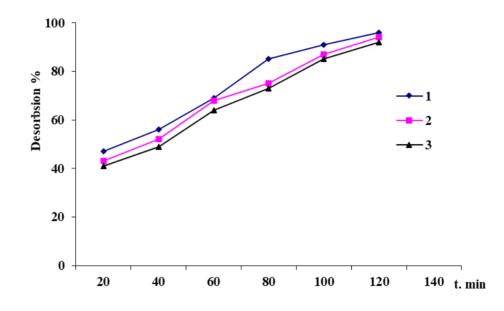


Fig 8. Dependence of desorption of vanadium from anion exchangers DGT+PEPA (1), DGT+M (2) and DGT+HPAN (3) on the duration of stirring.

The figure shows that the desorption rate is quite high, most of the vanadium is desorbed in the first 10 minutes of mixing. By selecting the conditions, especially the speed of stirring the solution, it is possible to achieve 100% desorption of vanadium.

To determine the limiting stage of desorption, the rate coefficients of internal and external diffusion were calculated and curves were plotted in coordinates $\ln(1-F) = \varphi(t)$ and $F = \varphi(\sqrt{t})$, as described above. It turned out that the rate of desorption is not limited by either external or internal diffusion. The process is also not in the mixed diffusion region, since the free term of the equation

$$y = a + bx, \tag{6}$$

where $y = \ln(1 - F)$, x = t

does not meet the specified limit of 0 to -0.49. Only the application of the equation of bimolecular chemical reaction gave a more or less constant rate constant:

$$\frac{dx}{dt} = k(a-x)(b-x) \tag{7}$$

from where
$$k = \frac{1}{t} \cdot \frac{1}{a-b} ln \frac{b(a-x)}{a(b-x)}$$
 (8)

The rate constant K was found to be equal for DGT + PEPA $(4,4\pm1)\cdot10^{-3}$, DGT+M $(5,2\pm1)\cdot10^{-3}$, and DGT+HPAN $(6,8\pm2)\cdot10^{-3}$. Thus, it can be concluded that the rate of desorption of vanadium is controlled by a bimolecular chemical reaction and decreases with decreasing vanadium concentration in the resins.

DISCUSSION AND CONCLUSION

The sorption of ions of molybdenum and vanadium was studied on the obtained anion exchangers, depending on the pH-medium, the ionic form of the ion exchanger, the concentration of the studied ions, the presence of competing ions, etc. It was established using potentiometric titration that the cations of molybdenum, vanadium and other nonferrous metals are absorbed by the anion exchangers mainly due to complexation and partly due to the formation of hydroxides.

The sorption and selective characteristics of the anion exchanger obtained on the basis of thiourea, epichlorohydrin in the presence of melamine (DGT+M) with respect to molybdenum ions were investigated depending on the pH medium, the ratio of the solid phase to the liquid, ionic form of the ion exchanger, the concentration of the solutions under study, and the presence of competing sulfate ions. The kinetics of the absorption of molybdenum ions have been investigated.

It was found that the intradiffusion mechanism of kinetics has a dominant effect on the sorption rate. The optimal conditions for the absorption of molybdenum ions by the DGT+M anionite were found and the possibility of using this anionite in the sorption of molybdenum ions in hydrometallurgy was shown.

REFERENCES

- [1] Tu YJ, Chan TS, Tu HW, Wang SL, You CF, & Chang CK. Rapid and efficient removal/recovery of molybdenum onto ZnFe2O4 nanoparticles. Chemosphere, 2016; 148: p 452-458. https://doi.org/10.1016/j.chemosphere.2016.01.054
- [2] Mostafa K, Kamran D. Removal of molybdenum using silver nanoparticles from water samples: particle swarm optimization-artificial neural network, J. Ind. Eng. Chem. 2014; 20(5): p 3014-3018. doi:10.1016/j.jiec.2013.11.036.
- [3] Bertoni FA, Medeot AC, Gonzalez JC, Sala LF, Bellu SE. Application of green seaweed biomass for Mo(VI) sorption from contaminated waters. Kinetic, thermodynamic and continuous sorption studies, J. Colloid Interface Sci. 2015; 446: p122–132.
- [4] Zeng J, Dong Z, Zhang Z, Liu Y, Preparation of a surface grafted imprinted ceramic membrane for selective separation of molybdate anion from water solutions, J. Hazard. Mater. 2017; 333:. p128–136.
- [5] Smedley P, Nicolli H, Macdonald D, Barros A, Tullio J. Hydrogeochemistry of arsenic and other inorganic constituents in groundwater from La Pampa, Argentina, Appl. Geochem. 2002; 17: p 259–284.
- [6] Attallah MF, Ahmed IM, Abd-Elhamid AI, Aly HF. Extraction of carrier-free 99Mo by ionic liquids from acid solutions: A model of Seaborgium (Sg) experiment, Appl. Radiat. Isot. 2019; 149: p 83–88.
- [7] Attallah MF, Moussa SI, Ahmed IM. Distribution coefficient properties of carrier free 99Mo as a homolog of Seaborgium (Sg) from some acid solutions using ion exchange resin, J. Molec. Liq. 2019; 277:p 323–329.
- [8] Cubova K, Semelova M, Nemec M, et al., Extraction of thallium and indium isotopes as the homologues of nihonium into the ionic liquids, J. Radioanal. Nucl. Chem. 2018; 318: p 2455–2461.
- [9] Ooe K, Attallah MF., Asai M, et al. Development of a new continuous dissolution apparatus with a hydrophobic membrane for superheavy element chemistry, J. Radioanal. Nucl. Chem. 2015; 303:p 1317–1320.
- [10] Toyoshima A, Ooe K, Miyashita S, et al. Chemical studies of Mo and W in preparation of a seaborgium (Sg) reduction experiment using MDG, FEC, and SISAK, J. Radioanal. Nucl. Chem. 2015; 303: p 1169–1172.
- [11] Valenzuela FR, Andrade JP, Sapag J, Tapia C, Basualto C. The solvent extraction separation of molybdenum and copper from acid leach residual solution of Chilean molybdenite concentrate miner. Eng. 1995.8, p 893–904.
- [12] Magyar MJ, Vanadium. U.S. Geological Survey Minerals Yearbook 2003.
- [13] Rao SRR. Resource Recovery and Recycling From Metallurgical Wastes. Elsevier, Amsterdam. 2011.
- [14] Syed S. Recovery of gold from secondary sources—a review. Hydrometallurgy 2012; 115: p 30–51.
- [15] Rivas BL, Maturana HA, Peric IM, Fierro AM. Modified poly(ethyleneimine) supports for vanadium, molybdenum, and rhenium removal. Polym. Bull. 1995; 35: p 337–343.
- [16] Soldi T, Pesavento M, Alberti G. Separation of vanadium(V) and -(IV) by sorption on animinodiacetic chelating resin. Anal. Chim. Acta 1996; 323: p 27–37.

- [17] Chen JH, Kao YY, Lin CH. Selective separation of vanadium from molybdenum using D2EHPA-immobilized Amberlite XAD-4 resin. Sep. Sci. Technol. 2003; 38: p 3827–3852.
- [18] Pyrzynska K, Wierzbicki T, Pre-concentration and separation of vanadium on Amberlite IRA-904 resin functionalized with porphyrin ligands. Anal. Chim. Acta 2005; 540: p 91–94.
- [19] Nomngongo PN, Ngila JC, Kamau JN, Msagati TAM, Moodley B. Preconcentration of molybdenum, antimony and vanadium in gasolsine samples using Dowex 1-x8 resin and their determination with inductively coupled plasma– optical emission spectrometry. Talanta 2013;110: p 153–159.
- [20] Kailasam V, Rosenberg E. Oxyanion removal and recovery using silica polyamine composites. Hydrometallurgy 2012; 129–130; p 97–104.
- [21] Li X, Liu R, Wu Sh, Liu J, Cai Sh, Chen D. Efficient removal of boron acid by N-methyl-D-glucamine functionalized silica-polyallylamine composites and its adsorption mechanism. J. Colloid Interface Sci. 2011; 361(1): p 232–237. doi: 10.1016/j.jcis.2011.05.036
- [22] Namasivayam C, Sureshkumar MV. Removal and recovery of molybdenum from aqueous solutions by adsorption onto surfactant-modified coir pith, a lignocellulosic polymer. CLEAN – Soil, Air, Water 2009; 37 (1): p 60–66.
- [23] Polowczyk I, Urbano BF, Rivas BL, Bryjak M, Kabay N. Equilibrium and kinetic study of chromium sorption on resins with quaternary ammonium and N-methyl-Dglucamine groups. Chem. Eng. J. 2016; 284: p 395–404.
- [24] Urbano BF, Rivas BL, Martinez F, Alexandratos SD. Equilibrium and kinetic study of arsenic sorption by water-insoluble nanocomposite resin of poly[N-(4- vinylbenzyl)-N-methyl-D-glucamine]-montmorillonite. Chem. Eng. J. 2012a; 193–194: p 21–30.
- [25] Urbano BF, Rivas BL, Martinez F, Alexandratos SD. Water-insoluble polymer– clay nanocomposite ion exchange resin based on N-methyl-D-glucamine ligand groups for arsenic removal. React. Funct. Polym. 2012b; 72(9): p 642–649.
- [26] Duranoglu D, Kaya IGB, Beker U, Senkal BF. Synthesis and adsorption properties of polymeric and polymer-based hybrid adsorbent for hexavalent chromium removal, Chem. Eng. J. 2012; 181-182: p 103-112. doi:10.1016/j.cej.2011.11.028
- [27] Xie M, Zeng L, Zhang Q et al. Synthesis and adsorption behavior of magnetic microspheres based on chitosan/organic rectorite for lowconcentration heavy metal removal, J. Alloys Compd. 2015; 647: p 892-905.
- [28] Chung TH, Pan HC, Lee WC, Preparation and application of magnetic poly(styrene-glycidyl methacrylate) microspheres, J. Magn. Magn. Mater. 2007; 311: p 36-40.
- [29] Horak D, Pollert E, Trchova M, Kovarova J. Magnetic poly(glycidyl methacrylate)-based microspheres prepared by suspension polymerization in the presence of modified La0.75Sr0.25MnO3 nanoparticles, Eur. Polym. J. 2009; 45: p 1009-1016. doi:10.1016/J.EURPOLYMJ.2008.12.030
- [30] Jovanovic SM, Nastasovic A, Jovanovic NN, Jeremic K, Savic Z. The influence of inert component composition on the porous structure of glycidyl methacrylate/ethylene glycol dimethacrylate copolymers, Die Angew. Makromol. Chem. 1994; 219 : p 161-168.
- [31] Miletic N, Vukovic Z, Nastasovic A, Loos K. Effect of Candida Antarctica lipase B immobilization on the porous structure of the carrier, Macromol. Biosci. 2011; 11: p 1537-1543.
- [32] Sandic ZP, Nastasovic AB, Jovic-Jovicic NP, Milutinovic-Nikolic AD, Jovanovic DM. Sorption of textile dye from aqueous solution by macroporous amino-functionalized copolymer, J. Appl. Polym. Sci. 2011; 121: p 234-242.
- [33] Hercigonja RV, Maksin DD, Nastasovic AB. et al. Adsorptive removal of technetium 99 using macroporous poly(- GMA-co-EGDMA) modified with diethylene triamine, J. Appl. Polym. Sci. 2012; 123: p 1273-1282.
- [34] Maksin DD, Nastasovic AB, Milutinovic-Nikolic AD. et al. Equilibrium and kinetics study on hexavalent chromium adsorption onto diethylene triamine grafted glycidyl methacrylate based copolymers, J. Hazard. Mater. 2012; 209-210: p 99-110.
- [35] Nastasovic A, Jovanovic S, Dordevic D, Onjia A, Jakovljevic D, Novakovic T. Metal sorption on macroporous poly(GMA-co-EGDMA) modified with ethylene diamine, React. Funct. Polym. 2004; 58: p 139-147.
- [36] Nastasovic A, Jakovljevic D, Sandic Z. et al. Amino-functionalized glycidyl methacrylate based macroporous copolymers as metal ion sorbents, in: M.I. Barroso (Ed.), React. Funct. Polym. Res. Adv, Nova Science Publishers Inc, New York, 2007; 79-112 pp.
- [37] Andronov EA, Lukacheva TM, Kuzin IA. Using the ion-exchange method for the separation of molybdenum from concentrated soda solutions. J. Appl. Chemistry, 1973; - T. VIII: p 1703-1706.
- [38] Plaksin IN, Tetargu SA. Hydrometallurgy using ion exchangers. Moscow: Metallurgy, 1964; p 176.
- [39] Laskorin BN, Ryabova EK. (RF). A.S. 509609 RF, IPC S 08 F 8/42. A method of obtaining complex-forming ion exchangers- 1924587; Stated 05/17/73; Publ. 04/05/1976; Bulletin 13. - C.2.
- [40] Kunin R, Meyers R. Ion exchange resins. -M,: IL., 1982; 207 p.
- [41] Gustafsson JP. Arsenate adsorption to soils: modeling the competition from humic substances. Geoderma 2006; 136: 320-330.
- [42] Dijkstra JJ, Meeussen JCL, Comans RNJ. Evaluation of a generic multi-sorption model for inorganic soil contaminants. Environ. Sci. Technol. 2009; 43: p 6196-6201.
- [43] Weng LP, Vega FA, van Riemsdijk WH. Competitive and synergistic effects in pH dependent phosphate adsorption in soils: LCD modeling. Environ. Sci. Technol. 2011; 45: p 8420-8428.
- [44] Groenenberg JE, Dijkstra JJ, Bonten LTC, de Vries W, Comans RNJ, Evaluation of the performance and limitations of empirical partitionrelations and process based multisurface models to predict trace element solubility in soils. Environ. Pollut. 2012; 166: p 98-107.
- [45] Gabler HE, Gluh K, Bahr A, Utermann J. Quantification of vanadium adsorption by German soils. J. Geochem. Explor. 2009; 103: p 37-44.
- [46] Sauve S, Hendershot W, Allen HE.. Solid-solution partitioning of metals in contaminated soils: dependence on pH, total metal burden, and organic matter.Environ. Sci. Technol. 2000; 34: p 1125-1131.
- [47] Gustafsson JP, Akram M, Tiberg C. Predicting sulphate adsorption/desorption in forest soils: evaluation of an extended Freundlich equation. Chemosphere 2015; 119: p 83-89.
- [48] Dzhalilov AT, Eshkurbonov FB, Beknazarov KhS. Sorption of ions of copper and other metals by the obtained anion exchangers. Journal of Plastic Mass. 2017; №9-10: p 20-24.
- [49] Zelikman AN. Molybdenum. M: Metallurgy, 1970; 150 p.
- [50] Saldadze KM. Ion-exchange high-molecular compounds / KM Saldadze, AB Pashkov, VS Titov. 3rd ed., Rev. and. add. M: Nauka, 1960; 71 p.

- [51] GOST 10896-84. Abstract and annotation. M: Publishing house of standards, 1984;. 10 p.
- [52] Polovinkina GM, Salazkina SN et al. Polymers based on polyethylenepolyamines metal sorbents. J. Appl. Chemistry. 1989; T. II: p. 337 - 341.
- [53] Askarov MA, Jalilov AT et al. A.S. 384374 Method of obtaining selective anionite /Publ. 1974; Bul. 27. C.2.
- [54] Kofman LYu, Vladimirova NYa, Gusakova NV. Neutralization of waste water from oxidative dehydrogenation of butane and n-butenes by a different exchange method. In Sat: Scientific. Proceedings of the Research Institute of Monomers for Synthetics. rubber. 1981; 4.: p. 47-51 and 356-359. (in Russian)

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